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Degradation Factors on Marine Fisheries in the Coastline of Dernah, East Libya

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ABSTRACT

This study has been conducted in the period between 2016 and 2019 on along the Derna coast. Severe decline has occurred in the case of marine fisheries in the area, where the Region was subjected to many pressures and crises, such as the subsequent closure of the fishing season due to war conditions and the widespread use of explosives in fishing (recording 896 accidents during the study period). The entry of foreign fishermen into the maritime fisheries sector without a license has resulted in a loss of productivity. The entry of foreign fishermen into the maritime fisheries sector without a license has resulted in a loss of productivity. The area reported a production of 5943 kilograms of fish in 2019 and decreased volume production from in 2015(25950 kilograms). and a significant decline in the number of fishing boats along the Derna coastline in the fishing sector. Decline from 200 different boats in 2015 to just 67 in 2019, and a significant decline in the number of fishing boats along the Derna coastline in the fishing sector from 200 different boats in 2015 to just 67 in 2019. resulting in very high fish and crustacean prices and a large number of local Fishers for other, more profitable occupations. In this study 83 native fish species, 4 endemic to the Mediterranean Sea, 20 non-indigenous of Lessepsian origin and 2 range expanding taxa from Gibraltar were recorded.

عوامل تدهور الصيد البحري في ساحل درنه - شرق ليبيا

أ.عزالنصر عاشور أبزيو

اجريت هذه الدراسة على طول ساحل درنه خلال الفترة من 2016 وحتى 2019 وقد ظهر تدهور شديد في حالة المصايد البحرية في المنطقة حيث تعرضت المنطقة للعديد من المؤثرات والازمات مثل الاغلاقات المتتالية للمصيد نظرا لظروف الحرب وانتشار استخدام المتفجرات في الصيد (تسجيل 896 حادثه خلال فترة الدراسة) ودخول عناصر اجنبية للعمل في مجال الصيد البحري دون ترخيص نتج عن ذلك انخيار في الانتاجية لتسجل المنطقة انتاج 5943 كيلو جرام سمك عام 2019م بانخفاض بنسبة 22.9% عن انتاج عام 2015م (25950 كيلوجرام) ونقصان شديد في اعداد المراكب العاملة في مجال الصيد على طول ساحل درنه لتتخلف من 200 مركب مختلفة الانواع عام 2015م الى 67 مركب فقط عام 2019م والامر الذي ادى الى ارتفاع شديد في اسعار الاسماك والقشريات وانتقال عدد كبير من الصيادين المحليين للعمل في مجالات اخرى اكثر ربحا. وقد نتج عن هذه الدراسة القائمة الاخيره لانواع الاسماك المسجلة في ساحل درنه ورصد للانواع المهاجرة والتي سجلت 83 نوع محلي , 4 متوطنه البحر المتوسط , 20 نوع مهاجر من قناة السويس و2 نوع امتد من مضيق جبل طارق على مدخل المتوسط من الاطلسي .

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INTRODUCTION

Libya, situated in North Africa, occupies the eastern segment of the Arabic Maghreb region, situated between latitudes 22° and 32° N and longitudes 10° and 25° E (Metz 1989; Otman and Karlberg, 2017), Libya encompasses an area of approximately 1,775,500 square kilometers, making it the fourth largest country in Africa (Karlberg and Otman 2007). The northern boundary of Libya is defined by the Mediterranean Sea, while it shares its western borders with Tunisia and Algeria, its eastern border with Egypt, and its southern borders with Sudan, Nigeria, and Chad (FAO 2005; Otman and Karlberg, 2007). Among Mediterranean nations, Libya boasts some of the largest, most intriguing, and least explored marine basin areas (Coll et al., 2010; 2012). Its coastline, extending roughly 2,000 kilometres along the Mediterranean, experiencing relatively minimal anthropogenic influence (Badalamenti et al., 2011), this region is currently facing an invasion of new species in its waters, alongside other areas of the eastern Mediterranean (Elbaraasi et al., 2013). The degradation of oceanic and marine ecosystems is significantly attributed to improper trading practices and various detrimental activities, which have largely been overlooked due to their limited direct impact on human populations (Mor, 2023). Yin et al. (2022) identified several critical factors contributing to the decline of fisheries, including overfishing, pollution, climate change, dam construction, developmental activities, illegal shipping operations, and the introduction of invasive species. The Mediterranean Sea, a largely enclosed marine area surrounded by the continents of Europe, Asia, and Africa, is connected to the Atlantic Ocean through the Strait of Gibraltar, which measures fifteen kilometers in width and has an average depth of 290 meters, with a maximum depth of 950 meters. Additionally, since the opening of the Suez Canal in 1869, the Mediterranean has had a link to the Red Sea. Research indicates that approximately 1,255 fish species have been recorded in the Northeast Atlantic and the Mediterranean Sea (EU, 2008), with 540 species specifically identified in the Mediterranean region. Among these, 362 species are found in coastal habitats, including 62 species that are endemic to the area (Tortonese, 1963). Given the clear evidence of regional speciation within the Mediterranean, it is unreasonable to assume uniformity in species composition across the entire sea (Whitehead et al., 1984).

Numerous investigations have been undertaken in the waters surrounding Libya. The initial study was

conducted by Vinciguera in 1881, who documented seventeen species while examining the ichthyofauna of the region. The early 20th century saw a rapid increase in the number of recognized species, as noted by Nini in 1914 and Tortonese in 1939. More comprehensive research was carried out in the latter half of the 20th century; for example, Aldebert and Pichot (1973) focused on certain flatfish species, while Duclerc (1973) concentrated on the Scorpaenidae family. Additional surveys produced checklists, such as one from the western region in 1972, which identified sixty-two species (Gorgy, 1972). In 1977, researchers had documented 39 species of cartilaginous fish and 185 species of bony fish. Zupanovic and El-Buni (1982) employed demersal fishing methods to assess the fish populations in Libyan waters, concluding that these populations exhibit a moderate potential for productivity. Their findings also indicated that Libya's ichthyofauna is largely linked to that of the eastern Mediterranean Sea, with a particular emphasis on the Levant Basin. In the eastern region of Libya, notably in Benghazi, a thorough survey of bony fish identified a total of 201 species distributed across seventy-one families and fifteen orders (Hassan and Silini, 1999). In 2007, Shakman and Kinzelbach examined the commercial fishery and species composition in Libya's coastal waters, providing researchers with updated information on both local and non-native fish species. Their findings included a new inventory of bony fish comprising approximately 104 species, contributing to a total of 304 taxa were recorded, including 271 species of natives, added by six endemic fish species to the Mediterranean, with 22 are of Lessepsian origin, and five are from Gibraltar. This updated data serves to monitor the various crises impacting fisheries along the coast of Derna in eastern Libya (Elbaraasi et al., 2013).

Blast fishing, also known as dynamite fishing, refers to the unlawful practice of employing explosives to incapacitate or kill fish populations for the purpose of easy collection. This method has been identified as a significant contributor to the degradation of numerous marine ecosystems worldwide (Fox and Caldwell 2006; England, 2014). It leads to decline in fish stocks and their habitats are destructed. Furthermore, the processes involved in ecosystem recovery remain poorly understood and can take considerable time (Ulaş et al.). In Libya, this illicit activity was prohibited by Law No. 14 in 1989, specifically in section 2, articles (7), (14), and (15). Nevertheless, the practice persists, particularly following the events of 2011, when a surge in fishing activities occurred amidst a climate of lawlessness,

exacerbated by the proliferation of weapons and explosives in the aftermath of the uprising (Buzaid et. al., 2022). Mor (2023) identified several critical factors contributing to the degradation of fish habitats and the fisheries sector, including overfishing, inadequate coastal development, population growth, unregulated shipping, and climate change, all of which stem from various detrimental human activities. This study aims to elucidate the significant impacts of fisheries degradation specifically along the Dernah shores on the eastern coast of Libya.

The main objective of this research is to examine the challenges associated with habitat degradation and the condition of fisheries along the coastline of Dernah. By analyzing the various factors contributing to degradation that affect both the fisheries sector and the habitats and diversity of fish, the study aims to identify essential strategies for mitigating the extent of degradation in future investigations.

RESEARCH METHODOLOGY

This study was performed between 2016 and 2019; in a survey aimed to find out the count and types of fishing units recorded and fishing gears used in each of four active landing sites Ras Al-hilal [32°52'58"N, 22°10'49"E](Buzaid, 2021), Dernah port [32° 16' 00" N, 22° 39' 12" E], Ras Attin [32°37' N-23°07' E] and Bumba Bay [32°25' N-23°06' E] (Reynolds et al., 1995 ; MBRC, 2005; Abu-Madinah, 2008), along the Derna coast in the eastern part of Libya (Fig. 1).

Primary and secondary data collection represent the two predominant methodologies employed in research studies (Korstjens & Moser, 2022). However, we collected important information about fishing vessels and fishing gear; according to method of Abziow, (2016) from local fishermen and fishermen's unions. that affected by many pressures and crises; such as the subsequent closure of the fishing season due to war crisis, and the widespread use of blast-fishing and the entry of un-licensed foreign fishermen into the fisheries sector (Buzaid et al., 2022). Taxonomically; fishes of this area were identified according to Whitehead et al., (1984), Fisher et al. (1987), Golani et al., (2006), Nelson (2006) and Ben Abdallah et al., (2009).

The Lessepsian fish species that recorded in the Mediterranean according to Azzurro et al. (2014) and Golani et al. (2017) as well.

Statically analyze

The collected data were examined and analyzed out by MS Excel 2010.



Figure (1). Map of Dernah coast, on the Mediterranean Sea, eastern Libya

RESULTS & DISCUSSION

Libya possesses the second largest continental shelf, encompassing approximately 65,000 km², and is home to some of the most productive fishing areas in the Mediterranean. Despite this potential, the fishing industry has remained largely underutilized until recent years, with less than 0.0025% of the Libyan population engaged in this sector (Serebetis 1952; Metz 1989; Reynolds et al. 1995; CMPE 2006; IUCN 2011). Nevertheless, the fishing industry continues to be a significant contributor to the national economy and supports a substantial export market (FAO, 2005).

Libya exhibits a significant degree of self-sufficiency in its fish supply, with an estimated per capita consumption of fresh fish products being relatively low at approximately 7 kg annually (FAO 2005). Notably, around 95% of the total fish catches are allocated for human consumption (European Commission 2009). According to Khalfallah et al. (2015), there has been a notable reconstruction of marine fisheries catches in Libya spanning from 1950 to 2010. However, since the outbreak of chaos and violence in 2011, the practice of

dynamite fishing has surged uncontrollably, exacerbated by the division of the country into two competing administrations that struggle to enforce regulations and manage oil ports. As a result, the safeguarding of fish stocks and the marine ecosystem has not been given adequate attention.

Experts have noted that the use of explosives, particularly those containing "gelatine," has devastating effects on fish populations, including the destruction of fish roe, larvae, and aquatic vegetation. Many of these explosives are homemade, leading to numerous fatalities and injuries among fishermen. Reports from Lana news agency in March highlighted the tragic deaths of three family members in Sirte due to blast fishing incidents (franch24.2018). This research indicates that artisanal fishing is crucial within the Libyan Exclusive Economic Zone (EEZ), while industrial fishing remains relatively insignificant. Since 2011, ongoing conflicts have resulted in a lack of oversight over fishing activities, which has likely led to a significant increase in illegal foreign fishing practices (Khalfallah et al., 2015).

State of fisheries sector

Fishing units of landing sites

Most of them were concentrated in the Derna port (44.8%) with 200 fishermen, Ras Attin site (23.9%) and (64) fishermen, Ras Al-hilal (16.4%) with 31 fishermen and (14.9%) with 25 of fishermen in Bumba Bay; in 320 fishermen as a total count (Table 1 & Figure 2& Figure3).

Compare to to the diversity in Benghazi, where more than this count of fishing workers were Libyans, Egyptians, Tunisians, Palestinians and Algerians and Africans as well such as in Benghazi port (MBRC, 2006 and Said, 2008); where the recent results were different, depending on the size, area, and the demographic difference in ports of Cyrenaica.

Table (1). Fishing units and fishermen in different fishing sites which operating along Derna coast (2016-2019).

Fishing sites	Count of fishing units (Average)	%	Count of fishermen
Derna Port	30	44.8	200
Ras El-Teen	16	23.9	64
Ras El-Helal	11	16.4	31
Gulf of Pomba	10	14.9	25
Total	67		320

Data obtained from General Authority for Developed of Fish Resources (Derna Office).

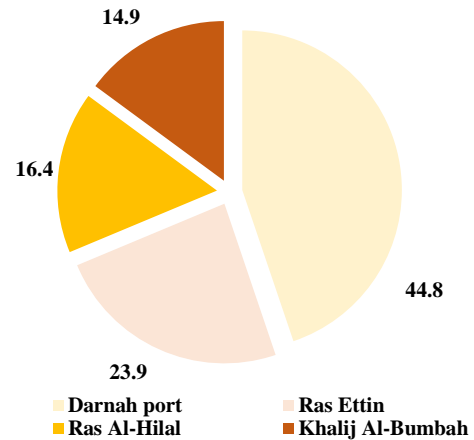


Figure (2): Counts and percentages of fishing units in different fishing sites on Derna coastline during (2016-2019).

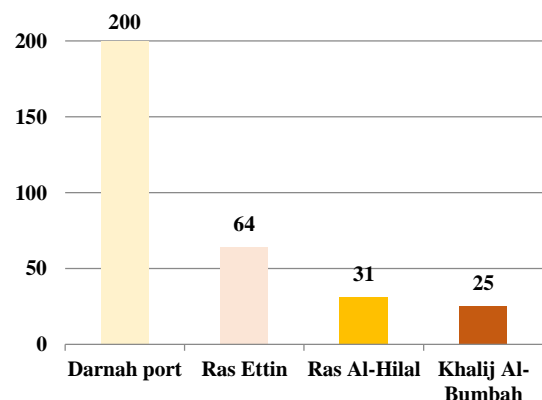


Figure (3): Counts of fishermen in different fishing sites on Derna coastline during (2016-2019).

Blast-fishing

A total of 896 accidents during the study period were observed in the coastlines of Derna port, Ras Al-hilal, Ras Attin and Bumba bay by 385 (43.0%), 233 (26.0%), 179 (20.0%) and 99 recoded accidents (11.0%) percent respectively (Figure 4). Blast fishing, an illicit method of fishing that employs explosives to easily eliminate fish stocks, has been prohibited worldwide, including in Libya, as stipulated in items No. (7), (14), and (15) of Law No. 14 of 1989. Nevertheless, this practice persists along Libyan coastlines, particularly due to its proliferation with little consequence, a situation exacerbated by the ongoing local conflicts that have inundated the country with weapons and explosives since 2011 (Buzaid et al., 2022). The regions affected by this destructive fishing technique require ages for recovery, and before fish populations can appear (UNEP, 2005).

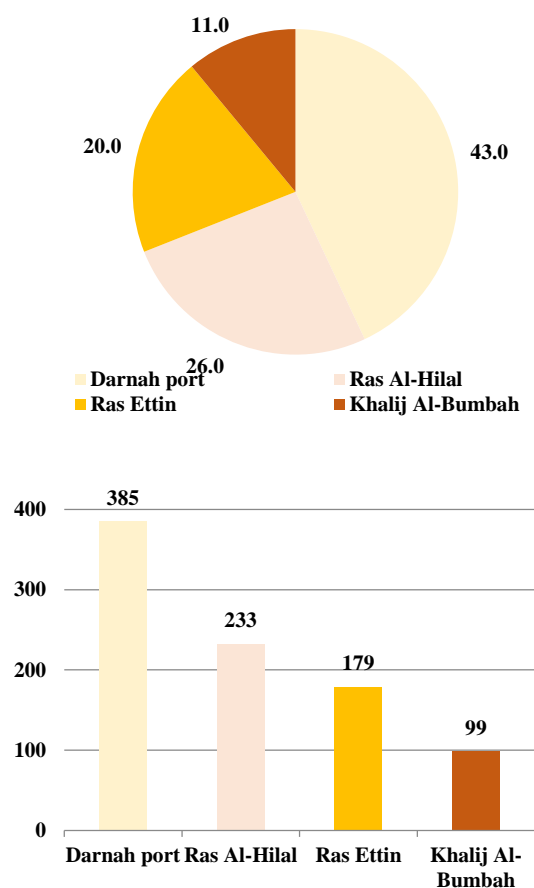


Figure (4) : The percentage of Blast fishing explosives use on the coast of Derna during the study period (2016-2019).

Data obtained from the organization of marine science and orgnasms nongovernmental organization (NGO) of Derna Office.



Figure (5) : The dynamites that used in Derna during the study period (2016-2019).

Source: Organization of marine sciences and organisms (OSMO, 2024).

Fishing fleets

In this work, about 67 fishing units were counted out in the active landing sites, the counted boats that were divided into four types: (i) Flouka by 65.7%, (ii) Mators (26.9%), (iii) Trawlers (4.5%) and (iv) Batah (2.9%) (Table 2 & Fig. 6). These pre-dominant percentages of fishing units are close to results of MBRC (2006) in ports of Benghazi, Abziew (2016) in Derna and Buzaid et al., (2024) in Tubruk. In contrast, a significant record of 32 units of a trawlers was capturing the bulk of the fishing production in Benghazi coasts (Buzaid, 2008), and as indicated by Lamboeuf et al. (2000), in Al-Tamimi and Ain al-Ghazala, Flouka units were used in less than 50 m depth, while Batah boats are used at depths of up to 5 m depth.

Table (2). Types and percentage of fishing boats which operating along Derna coast during 2016-2019.

Types of fishing boats	The average no. of boats	%
Flouka	44	65.7
Mator	18	26.9
Trawling	3	4.5

Batah	2	2.9
Total	67	

Data obtained from The General Authority for the Developed of Fish Resources (Derna Office).

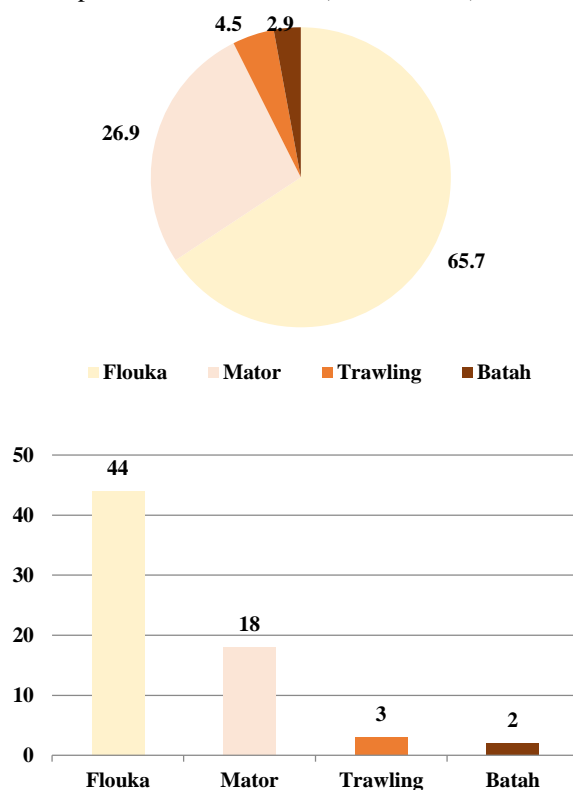


Figure (6): Counts and percentages of fishing units in different fishing sites on Derna coastline during (2016-2019).

Fish production yield:

In 2007, the total fish production reached 39,314 kg, which saw a significant rise to 160,456 kg by 2010. However, this figure plummeted to its lowest point of 14,759 kg in 2011. Subsequently, production rebounded to 148,315 kg in 2013, only to decline again to 25,950 kg in 2015 (Abziow, 2016). The downward trend continued, culminating in a minimal output of 1,480 kg in 2017, before experiencing a modest increase to 5,943 kg in 2019 (Table 3 & Figure 4).

In General; the marine fisheries in the study area have experienced a significant decline, attributed to various environmental pressures and crises. Notably, the fishing season was interrupted due to wartime conditions, and the prevalent use of explosives in fishing led to 896 recorded incidents during the study period. Additionally, the unauthorized entry of foreign fishing vessels into the Libyan maritime fisheries sector has contributed to a decrease in productivity. In 2019, fish production was reported at 5,943 kilograms, reflecting a 22.9% reduction

from the 25,950 kilograms produced in 2015. Furthermore, there has been a marked decrease in the number of fishing vessels operating along the Derna coastline, dropping from 200 boats in 2015 to merely 67 by 2019. This decline has resulted in soaring prices for fish and crustaceans, prompting many local fishermen to seek alternative, more lucrative employment opportunities.



Figure (7): The fishing units in Derna landing sites
Source: Organization of marine sciences and organisms (OSMO, 2024).

The research revealed the presence of 83 native species, among which 4 are endemic to the Mediterranean region, alongside 20 non-native species originating from the Lessepsian migration, and 2 taxa that are extending their distribution from Gibraltar. A significant number of these marine alien species exhibit highly invasive traits, thereby imposing substantial ecological stress on both marine and estuarine ecosystems (Sala et al., 2011). The main routes through which species are introduced into the Mediterranean Sea include the Suez Canal, maritime activities (such as the discharge of ballast water and sediments, anchoring practices, and biofouling), aquaculture (which includes both marine and brackish water species), and the commerce of live marine

organisms (covering aquarium trade, fishing bait, and seafood). This observation is consistent with recent evaluations of introduction pathways at the Pan-European scale (Katsanevakis et al., 2013a).

Table (3). Total annual fish production for Derna coast during 2016- 2019.

Year	Total fish production (Kg)
2016	2258
2017	1486
2018	*
2019	5943
Total catch	9687

* Fish productivity was not recorded due to security conditions. Data obtained from General Authority for Developed of Fish Resources (Derna Office).

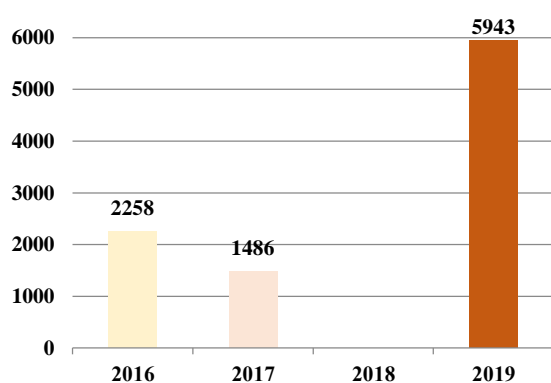


Figure (8) : Annual fish landing of Derna coastline (2016 - 2019)

Ichthyofaunal wealth

Talking about 83 native, 4 endemic to the Mediterranean Sea, 20 non indigenous of Lessepsian origin and 2 range expanding taxa from Gibraltar, which observed here. The most commercially-important families of bony fishes in Derna coastline; include Anguillids, Clupeids, Serranids, Sparids, Mugilids and so on.

About the recorded Red Sea migrants; take examples: *Herklotsichthys punctatus*, *Etrumeus golanii*, *Saurida lessepsianus*, *Atherinomorus forskalii*, *Hemiramphus far*, *Parexocoetus mento*, *Sargocentron rubrum*, *Fistularia commersonii*, *Alepes djedaba*, *Crenidens crenidens*, *Upeneus moluccensis*, *Upeneus pori*, *Liza carinata*, *Siganus luridus*, *Siganus rivulatus*, *Scomberomorus commerson*, *Lagocephalus sceleratus* and *Lagocephalus suezensis*. Whereas the stretched species from the

Atlantic Ocean via Gibraltar were two species: *Seriola fasciata* and *Seriola rivoliana*.

Looking below to the checklist of bony fishes species which were collected from the different sites in Derna coast (2016-2019) [N = Native; E = Endemic of the Mediterranean Sea; A = Atlantic Ocean origin; L = Lessepsian species].

Anguillida

Anguilla anguilla (Linnaeus, 1758) European eel (N)

Muraenidae

Muraena helena (Linnaeus, 1758)

Mediterranean moray (N)

Conger conger (Linnaeus, 1758)

European conger (N)

Clupeidae

Alosa alosa (Linnaeus, 1758)

Allis shad (N)

Herklotsichthys punctatus (Rüppell, 1837)

Spotback herring (L)

Sardina pilchardus (Walbaum, 1792)

European pilchard (N)

Sardinella aurita (Valentines, 1847) Round sardinella (N)

Sardinella maderensis (Lowe, 1838) Madeiran sardinella (N)

Sprattus sprattus (Linnaeus, 1758) European sprat (N)

Etrumeus golanii (DiBattista, Randall & Bowen, 2012)

Round herring (L)

Engraulidae

Engraulis encrasicolus (Linnaeus, 1758) European anchovy (N)

Gonostomatidae

Cyclothone braueri (Jespersen & Tåning, 1926) Garrick (N)

Cyclothone pygmaea (Jespersen & Tåning, 1926) pygmy bristlemouth (E)

Stomiidae

Stomias boa (Risso, 1810) Boa dragonfish (N)

Synodontidae

Synodus saurus (Linnaeus, 1758) Atlantic lizardfish (N)
Saurida lessepsianus (Russell, Golani & Tikochinski, 2015) (L)

Arctozenus risso (Bonaparte, 1840) Spotted barracudina (N)

Lestidiops jayakari (Boulenger, 1889) Pacific barracudina (N)

Paralepis coregonoides (Risso, 1820) Sharpchin barracudina (N)

Phycidae

Phycis blennoides (Brünnich, 1768) Greater forkbeard (N)

Phycis phycis (Linnaeus, 1766) Fork beard(N)

Atherinidae

Atherina boyeri (Risso, 1810) Big-scale sand smelt (N)

Atherina hepsetus (Linnaeus, 1758) Mediterranean sand smelt (N)

Atherinomorus forskalii (Rüppell, 1838) hardyhead silverside (L)

Belonidae

Belone belone (Linnaeus, 1761) Garfish (N)

Hemiramphidae

Hemiramphus far (Forsskål, 1775) Black-barred halfbeak (L)

Hyporhamphus picarti (Valenciennes, 1847) African halfbeak (N)

Exocoetidae

Cheilopogon heterurus (Rafinesque, 1810)

Mediterranean flyingfish (N)

Parexocoetus mento (Valenciennes, 1847) African sailfin flyingfish (L)

Holocentridae

Sargocentron rubrum (Forsskål, 1775) Redcoat (L)

Fistulariidae

Fistularia commersonii (Rüppell, 1838) Blue-spotted Cornetfish (L)

Syngnathidae

Hippocampus guttulatus (Cuvier, 1829) Long-snouted seahorse (N)

Hippocampus hippocampus (Linnaeus, 1758) Short snouted seahorse (N)

Scorpaenidae

Scorpaena elongata (Cadenat, 1943) Slender rockfish (N)

Scorpaena loppei (Cadenat, 1943) Cadenat's rockfish (N)

Triglidae

Chelidonichthys cuculus (Linnaeus, 1758) Red gurnard (N)

Lepidotrigla cavillone (Lacepède, 1801) Large-scaled gurnard (N)

Lepidotrigla dieuzeidei (Blanc & Hureau, 1973) Spiny gurnard (N)

Trigla lyra (Linnaeus, 1758) Piper gurnard (N)

Moronidae

Dicentrarchus labrax (Linnaeus, 1758) European seabass(N)

Dicentrarchus punctatus (Bloch, 1792) Spotted seabass (N)

Serranidae

Epinephelus aeneus (Geoffroy Saint-Hilaire, 1817) White grouper (N)

Epinephelus caninus (Valenciennes, 1843) Dogtooth grouper (N)

Epinephelus fasciatus (Forsskål, 1775) Blacktip grouper (N)

Epinephelus marginatus (Lowe, 1834) Dusky grouper (N)

Serranus cabrilla (Linnaeus, 1758) Comber (N)

Serranus hepatus (Linnaeus, 1758) Brown comber (N)

Serranus scriba (Linnaeus, 1758) Painted comber (N)

Apogonidae

Apogon imberbis (Linnaeus, 1758) Cardinalfish (N)

Carangidae

Alepes djedaba (Forsskål, 1775) Shrimp scads (L)

Caranx rhonchus (Geoffroy Saint-Hilaire, 1817) false scad (N)

Lichia amia (Linnaeus, 1758) Leerfish (N)

Seriola fasciata (Bloch, 1793) lesser amberjack (A)

Seriola rivoliana (Valenciennes, 1833) Longfin yellowtail (A)

Trachurus mediterraneus (Steindachner, 1868)

Mediterranean horse mackerel (N)

Sparidae

Boops boops (Linnaeus, 1758) Bogue (N)

Crenidens crenidens (Forsskål, 1775) Karanteen seabream (L)

Dentex dentex (Linnaeus, 1758) Common dentex (N)

Diplodus annularis (Linnaeus, 1758) Annular seabream (N)

Diplodus cervinus (Lowe, 1838) Zebra seabream (N)

Diplodus puntazzo (Walbaum, 1792) Sharp snout seabream (N)

Diplodus sargus (Linnaeus, 1758) White Seabream (N)

Diplodus vulgaris (Geoffroy Saint-Hilaire, 1817)

Common two-banded seabream (N)

Lithognathus mormyrus (Linnaeus, 1758) Sand
steenbras (N)

Oblada melanura (Linnaeus, 1758) Saddled bream (N)

Pagellus acarne (Risso, 1827) Axillary seabream (N)

Pagellus bellottii (Steindachner, 1882) Red pandora (N)

Pagellus bogaraveo (Brünnich, 1768) Blackspot
seabream (N)

Pagellus erythrinus (Linnaeus, 1758) Common pandora
(N)

Pagrus auriga (Valenciennes, 1843) Redbanded
seabream (N)

Pagrus caeruleostictus (Valenciennes, 1830)

Bluespotted seabream (N)

Pagrus pagrus (Linnaeus, 1758) Red porgy (N)

Sarpa salpa (Linnaeus, 1758) Salem (N)

Sparus aurata (Linnaeus, 1758) Gilthead seabream (N)

Spondylisoma cantharus (Linnaeus, 1758) Black
seabream (N)

Sciaenidae

Argyrosomus regius (Asso, 1801) Meager (N)

Sciaena umbra (Linnaeus, 1758) Brown meager
(N)

Umbrina cirrosa (Linnaeus, 1758) shi drum (N)

Umbrina canariensis (Valenciennes, 1843) Canary drum
(N)

Mugilidae

Chelon labrosus (Risso, 1827) Thicklip grey mullet (N)

Liza aurata (Risso, 1810) Golden grey mullet
(N)

Liza carinata (Valenciennes, 1836) Keeled mullet (L)

Liza ramada (Risso, 1827) Thinlip grey mullet
(N)

Liza saliens (Risso, 1810) Leaping mullet
(N)

Mugil cephalus (Linnaeus, 1758) Flathead grey mullet
(N)

Mullidae

Mullus barbatus (Linnaeus, 1758) Red mullet
(N)

Mullus surmuletus (Linnaeus, 1758) Surmullet
(N)

Upeneus moluccensis (Bleeker, 1855) Goldband goatfish
(L)

Upeneus pori (Ben-Tuvia & Golani, 1989) Por's
goatfish (L)

Labridae

Coris julis (Linnaeus, 1758) Mediterranean rainbow
wrasse (N)

Symphodus doderleini (Jordan, 1890) wrasse
(E)

Blenniidae

Blennius ocellaris (Linnaeus, 1758) Butterfly blenny
(N)

Salaria basilisca (Valenciennes, 1836) (E)

Gobiidae

Deltentosteus quadrimaculatus (Valenciennes, 1837)

Four-spotted goby (N)

Gobius niger (Linnaeus, 1758) Black goby (N)

Pomatoschistus tortonesei (Miller, 1969) Tortonese's
goby (E)

Siganidae

Siganus luridus (Rüppell, 1829) Dusky spine foot (L)

Siganus rivulatus (Forsskål & Niebuhr, 1775) Marbled
spinefoot (L)

Sphyraenidae

Sphyraena sphyraena (Linnaeus, 1758) European
barracuda (N)

Sphyraena viridensis (Cuvier, 1829) Yellowmouth
barracuda (N)

Sphyraena chrysotaenia (Klunzinger, 1884) Obtuse
barracuda (L)

Sphyraena flavicauda (Rüppell, 1838) Yellowtail
barracuda (L)

Scombridae

Scomberomorus commerson (Lacepède, 1800) Narrow-
barred Spanish mackerel (L)

Thunnus alalunga (Bonnaterre, 1788) Albacore (N)

Soleidae

Microchirus ocellatus (Linnaeus, 1758) Foureyed sole
(N)

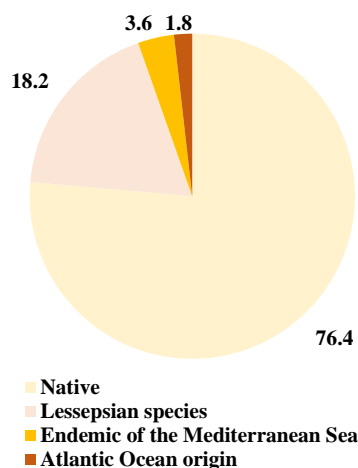
Solea aegyptiaca (Chabanaud, 1927) Egyptian sole (N)

Solea solea (Linnaeus, 1758) Common sole (N)

Tetraodontidae

Lagocephalus lagocephalus (Linnaeus, 1758) Oceanic
puffer (N)

Lagocephalus sceleratus (Gmelin, 1789) Silver-cheeked
toadfish (L)



Lagocephalus suezensis (Clark & Gohar, 1953) Suez pufferfish

Since its opening in 1869, the Suez Canal has facilitated the entry of at least 30 marine species from the Indo-Pacific region into the Mediterranean, a phenomenon known as Lessepsian migration, named in honor of Ferdinand de Lesseps (Por, 1978; Ben-Abdalla & Al-Turkey, 2005). Notably, around 65 fish species have been recorded in the Mediterranean, and the list of newly arrived species continues to expand (Golani, 2006).

The introduction of various species has resulted in considerable changes to the species composition within Mediterranean communities, leading to the formation of mixed Red-Mediterranean communities (Fishelson, 2000). Although it is clear that Lessepsian fish migrant species have had a significant impact on the eastern Mediterranean ecosystem, a thorough investigation to assess this influence is currently lacking. The Mediterranean biota is progressively transitioning towards a state characterized by 'tropicalization,' largely driven by the influx of numerous alien species with tropical origins and affinities (Bianchi & Mori, 2003). Libya, located in the Central Mediterranean Basin as outlined by the Marine Strategy Framework Directive (MSFD) (EU, 2008), encounters difficulties in obtaining comprehensive data on alien species, which are frequently limited and fragmented. This limitation is primarily due to the lack of long-term monitoring programs within national frameworks that address alien species. An inventory of exotic fish has recorded 16 introduced species (Ben-Abdallah et al., 2005), which has been further updated by Shakman & Kinzelbach (2007c), Milazzo et al. (2012), and Sghaier et al. (2013). All species identified in the current study are included in the checklist compiled by Hassan and El-Silni (1999) as well as the list by El-Baraasi et al. (2013).

4. CONCLUSION

The current investigation reveals a significant deterioration in marine fisheries along the Derna coast, attributed to various environmental pressures and crises. Notably, fishing activities were suspended in 2017 and 2018 due to wartime conditions and the prevalent use of explosives for fishing. The influx of unlicensed foreign fishermen has further exacerbated productivity losses, with blast fishing playing a critical role in this decline. However, a modest recovery was noted in 2019, with fish production reaching 5,943 kilograms, following a substantial decrease of 22.9% from 25,950 kilograms in 2015 to no production in 2018. This decline was primarily due to a drastic reduction in the number of operational vessels along the Derna coastline, which fell from 200 boats in 2015 to merely 67 in 2019.

Consequently, the prices of finfish and shellfish surged, prompting many local fishermen to seek alternative, more lucrative employment opportunities. Additionally, this study identified 83 native fish species, including 4 endemics to the Mediterranean, 20 invasive species from the Red Sea, and two expanding taxa from Gibraltar, highlighting the biodiversity present in the region. Talking about the blast-fishing; to shorten the recovery period and enhance the restoration process, extensive human intervention is essential, particularly when recovery is deemed necessary. Initially, it is crucial to formulate a comprehensive fisheries management action plan. This plan should encompass either a temporary or permanent prohibition or suspension of fishing activities in the impacted regions, with the aim of restoring both the ecological and economic value of the coastal ecosystem. The involvement of governmental bodies, organizations, and local authorities—especially fishermen's unions, coast guards, marine resource offices, and civil society organizations—is vital to ensure effective oversight and enforcement of regulations in areas where violations are prevalent.

In the end; the paramount objective is to enhance "community awareness" regarding the dangers and detrimental influences affecting marine ecosystems and fisheries within the region. This approach involves leveraging the religious, moral, and tribal values of the local community, as these elements are often more effective and resonant than traditional legal measures in addressing these issues.

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